# Seasonal Dynamics of the Zooplankton Community in the Temperate Eutrophic Çaygören Reservoir (Balıkesir), Turkey Related to Certain Physicochemical Parameters of Water 

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#### Abstract

Seasonal dynamics in the zooplankton community and its relationships with certain physicochemical parameters (water temperature, Secchi disk transparency, conductivity, nitrate-nitrogen $\left(\mathrm{NO}_{3}-\mathrm{N}\right)$, phosphate $\left(\mathrm{PO}_{4}\right), \mathrm{pH}$, oxidation-reduction potential (ORP) and chlorophyll-a) were studied in the eutrophic Çaygören Reservoir, Turkey from February 2007 to March 2008. Cyclops vicinus was dominant throughout the year; Acanthocyclops robustus was dominant in fall and summer; Asplanchna priodonta was dominant in fall, winter and spring at all stations. Canonical correspondence analysis (CCA) showed that $A$. robustus had high correlations to conductivity, C . vicinus to $\mathrm{PO}_{4}$ and $\mathrm{NO}_{3}-\mathrm{N}$ and Metacyclops gracilis to water temperature and chl- $a$. The number of individual were significantly different among the seasons and stations ( $\mathrm{P}<0.05$ ). The number of species were significantly different among the seasons ( $\mathrm{P}<0.05$ ), but not among the stations ( $\mathrm{P}>0.05$ ). Zooplankton community in the Çaygören Reservoir underwent changes in species composition from small-bodied


## Introduction

The functioning of aquatic ecosystems depends on the biological diversity of the system (Veerendra, Thirumala, Manjunatha, \& Aravinda, 2012). Zooplankton community is an important biotic component of aquatic ecosystems, as it occupies a critical link between primary producers and consumers and contributes to the recycling of nutrients (Hood \& Sterner, 2010).

The major freshwater zooplankton groups include Cladocera, Copepoda and Rotifera. Cladocerans have the ability to survive in extreme conditions (Sarma, Nadini, \& Gulati, 2005). Copepods are important prey items for the young fish and they feed on intermediate consumers, mainly ciliates (Sommer et al., 2001; Feuchtmayr, Zollner, Santer B, Sommer, \& Grey, 2004).

Many rotifer species are indicators of water quality in inland waters (Rajashekhar, Vijaykumar, \& Parveen, 2009).

The distribution and diversity of zooplankton in aquatic ecosystems depend mainly on the physicochemical properties of the water column (Barnett \& Beisner, 2007). The temporal variations in zooplankton community may depend on changes in the availability of edible phytoplankton which often vary depending on the physical processes and nutrient availability in the water bodies (Sarmento et al., 2008).

An understanding of the community patterns will better allow us to predict the dynamics of zooplankton community structure under future scenarios of anthropogenically induced changes in lakes and reservoirs. Lately the zooplanktonic organisms of reservoirs in Turkey have attracted the attention of
many scientists (Baykal, Salman, \& Açıkgöz, 2006; Kaya \& Altındağ, 2007; Bozkurt \& Göksu, 2010; Buyurgan, Altındağ, \& Kaya, 2010; Gökçe \& Özhan, 2011; Apaydın Yağcı \& Ustaoğlu, 2012; Bozkurt \& Akın, 2012; Gökçe \& Turhan, 2014; Saler, Alpaslan, Karakaya, \& Gündüz, 2017).

There are still reservoirs in Turkey that its zooplanktonic organisms have not been studied yet. In this study, we aimed to assess the seasonal dynamics of zooplankton community in relation to certain physicochemical and biological parameters in the eutrophic temperate Çaygören Reservoir, Balıkesir, Turkey.

## Materials and Methods

## Study Area

In Turkey, the most important problem for agriculture is the lack of irrigation water during summer due to its dry and hot Mediterranean climate. One of the effective solutions for this problem is building dams on the running waters. The Çaygören Reservoir was built in 1971 for the purpose of irrigating the Sındırgı and Bigadiç plains. It is also used for power generation. Sport fishing in the reservoir is also popular. Therefore, the reservoir is important to the local and regional economic and ecological sustainability (Arslan \& Ergül, 2014).

The Çaygören Reservoir is located at $39^{\circ} 17{ }^{\prime} 24^{\prime \prime} \mathrm{N}$ and $28^{\circ} 19^{\prime} 16^{\prime \prime} \mathrm{E}, 55 \mathrm{~km}$ southeast of Balıkesir, Turkey (Figure 1). It lies at 273 m above the sea level and has a maximum depth of 28 m , a length of 4.6 km and a surface area of $9 \mathrm{~km}^{2}$. The reservoir is fed by the Simav

Stream (State Water Works, 2017).

## Sampling Procedure and Analysis

Sampling was started in February 2007 and ended in March 2008. Zooplankton was sampled with vertical net hauls using a 0.30 m diameter net with $60 \mu \mathrm{~m}$ mesh size at three stations. Vertical tows were carried out from the bottom to the surface.

A total of 100 L water was filtered through the net for a composite sample. The volume of filtered water was calculated by multiplying the area of the mouth of the net by the length (depth) through which the net was towed. A calibrated meter in the net corrected for back flushing during vertical hauls. Samples were fixed and preserved with $4 \%$ formaldehyde in 500 ml plastic bottles immediately after collection (APHA, 1995).

Zooplankton specimens were identified and counted under an inverted microscope. Counting was done in Petri dishes from 4 ml sub-samples. A minimum of 200 individuals were quantified per replicate, and the final density was converted to individuals per cubic meter. The zooplankton taxa were identified using the common taxonomic keys (Dussart, 1968; Ruttner-Kolisko, 1974; Harding \& Smith, 1974; Kiefer, 1978; Koste, 1978; Stemberger, 1979; Negrea, 1983; Amoros, 1984; Koste, 2000; Nogrady \& Segers, 2002; Benzie, 2005).

Measurements of water temperature, conductivity, pH , oxidation-reduction potential (ORP) and chlorophyll-a (chl-a) were taken using a YSI water quality multi-probe. Water transparency was measured using a Secchi disk. Phosphate $\left(\mathrm{PO}_{4}\right)$ and nitrate-


Figure 1. The map of the Çaygören Reservoir and the location of sampling stations.
nitrogen $\left(\mathrm{NO}_{3}-\mathrm{N}\right)$ concentrations were determined spectrophotometrically on filtered water (APHA, 1995).

An ANOVA test was carried out to test the spatial and temporal variations among stations and seasons using SPSS (ver. 11.0) software. Canonical Correspondence Analysis (CCA) was applied to the data to determine the relationships between the dominant zooplankton species and water temperature, conductivity, Secchi disk transparency, pH , oxidationreduction potential, $\mathrm{PO}_{4}, \mathrm{NO}_{3}-\mathrm{N}$ and chl-a using the CANOCO program (ter Braak \& Smilauer, 2002).

## Results

Chl- $a$ ranged from $\left.2.2 \mu \mathrm{~g}\right|^{-1}$ to $\left.30.0 \mu \mathrm{~g}\right|^{-1}$ at the first station, from $1.9 \mu \mathrm{gl}^{-1}$ to $37 \mu \mathrm{gl}^{-1}$ at the second station and from $\left.1.4 \mu \mathrm{~g}\right|^{-1}$ to $37 \mu \mathrm{gl}{ }^{-1}$ at the third station (Table 1). Secchi disk depth ranged from 0.3 m to 1.5 m at the first station, from 0.6 m to 1.95 m at the second station and from 0.6 m to 1.98 m at the third station (Table 1).

Water temperature ranged from $5^{\circ} \mathrm{C}$ to $26^{\circ} \mathrm{C}$ at the first station, from $4{ }^{\circ} \mathrm{C}$ to $26^{\circ} \mathrm{C}$ at the second station and from $5^{\circ} \mathrm{C}$ to $25.87{ }^{\circ} \mathrm{C}$ at the third station (Table 1). Conductivity ranged from $0.34 \mathrm{mScm}^{-1}$ to $0.63 \mathrm{mScm}^{-1}$ at the first station, from $0.33 \mathrm{mScm}^{-1}$ to $0.53 \mathrm{mScm}^{-1}$ at the second station and from 0.32 $\mathrm{mScm}^{-1}$ to $0.88 \mathrm{mScm}^{-1}$ at the third station (Table 1).
pH ranged from 8 to 11.6 at the first station, from 8.2 to 11 at the second station and from 8.4 to 11.1 at the third station (Table 1). Phosphate $\left(\mathrm{PO}_{4}\right)$ concentrations ranged from $0.003 \mathrm{mgl}^{-1}$ to $0.006 \mathrm{mgl}^{-1}$ at the first station, from $0.001 \mathrm{mg} \mathrm{l}^{-1}$ to $0.003 \mathrm{mgl}^{-1}$ at the second station and from $0.001 \mathrm{mgl}^{-1}$ to $0.006 \mathrm{mg}^{-1}$ at the third station (Table 1). Nitrate-nitrogen ( $\mathrm{NO}_{3}-\mathrm{N}$ ) concentrations ranged from $0.08 \mathrm{mgl}^{-1}$ to $0.25 \mathrm{mg}^{-1}$ at the first station, from $0.007 \mathrm{mgl}^{-1}$ to $1.6 \mathrm{mgl}^{-1}$ at the second station and from $0.007 \mathrm{mgl}^{-1}$ to $0.3 \mathrm{mgl}^{-1}$ at the third station (Table 1).

During the study, a total number of 9 rotifers, 7
cladocerans and 4 copepods were collected. At the first station, the most dominant species, Asplanchna priodonta (Rotifera), reached its peak abundance in November 2007 and Metacyclops gracilis (Copepoda) in July 2007. At the second station, the most dominant species, A. priodonta (Rotifera), reached its peak abundance in November 2007, Trichocerca capucina (Rotifera) in August 2007, Daphnia galeata (Cladocera) in November 2007 and M. gracilis (Copepoda) in February 2007. At the third station, the most dominant species, Filinia longiseta (Rotifera), reached its peak abundance in in September 2007 and Cyclops vicinus (Copepoda) in February 2008.

At the first station, 7 species of Rotifera, 2 Cladocera and 4 Copepoda were collected. The lowest number of species (1) was collected in February 2007 and 2008 and March 2008, the highest number (8) of species was collected in September 2007 (Table 2). At the second station, 8 species of Rotifera, 7 Cladocera and 3 Copepoda were collected (Table 3). The lowest number of species (4) was collected in December 2007 and March 2008 and the highest number (12) was collected in September 2007. At the third station, 7 species of Rotifera, 6 Cladocera and 4 Copepoda were collected (Table 4). The species number was at the lowest level (5) in February, March, May and December 2007 and in January and February 2008 and it reached at the highest level (11) in August and September 2007.

The average annual abundance of zooplankton for the Çaygören Reservoir was 17326.71 ind. $\mathrm{m}^{-3}$, the nauplius larvae of Cladocera and Copepoda were not included in counting because of the identification difficulties. At the first station, the lowest number of individuals ( 674 ind. $\mathrm{m}^{-3}$ ) was collected in February 2007 and the highest number ( 118569 ind. $\mathrm{m}^{-3}$ ) was collected in September 2007 (Table 2). At the second station, the lowest number of individuals ( 3529 ind. $\mathrm{m}^{-3}$ ) was collected in March 2007 and the highest number ( 68751 ind. $\mathrm{m}^{-3}$ ) was collected in August 2007 (Table 3).

Table 1. Minimum, maximum, Mean and Standard deviation physical and chemical water characteristics for water quality parameters in the Çaygören Reservoir

|  | Station 1 |  |  |  | Station 2 |  |  |  | Station 3 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Var. | Min. | Max. | Mean | Std. D. | Min. | Max. | Mean | Std. D. | Min. | Max. | Mean | Std. D. |
| Chl-a $\left(\left.\mu\right\|^{-1}\right)$ | 2.2 | 30.0 | 14.3 | 9.8 | 1.90 | 37.0 | 12.0 | 10.9 | 1.40 | 37.0 | 10.08 | 9.33 |
| Secch <br> (m) | 0.4 | 1.5 | 0.85 | 0.33 | 0.6 | 1.95 | 1.10 | 0.47 | 0.6 | 1.98 | 1.21 | 0.48 |
| Temp. $\left({ }^{\circ} \mathrm{C}\right)$ | 5.0 | 26.6 | 14.20 | 7.07 | 4.0 | 26.0 | 13.5 | 6.63 | 5.00 | 25.87 | 12.82 | 5.90 |
| Cond. (mScm ${ }^{-1}$ ) | 0.34 | 0.63 | . 44 | . 08 | . 33 | . 53 | . 44 | . 08 | . 32 | . 88 | . 45 | . 09 |
| pH | 8.9 | 11.0 | 9.84 | . 61 | 8.2 | 11.0 | 9.69 | . 66 | 8.40 | 11.1 | 9.72 | . 57 |
| $\begin{aligned} & \mathrm{PO}_{4} \\ & \left(\mathrm{mgl}^{-1}\right) \end{aligned}$ | 0.003 | 0.06 | . 03 | . 005 | . 001 | . 03 | . 022 | . 007 | . 001 | . 06 | . 023 | . 011 |
| $\begin{aligned} & \mathrm{NO}_{3} \\ & \left(\mathrm{mgl}^{-1}\right) \end{aligned}$ | . 08 | . 25 | . 18 | . 052 | . 07 | 1.6 | . 65 | . 27 | . 07 | . 30 | . 19 | . 052 |

At the third station, the lowest number of individuals ( 5952 ind. $\mathrm{m}^{-3}$ ) was collected in March 2007 and the highest number ( 74034 ind. $\mathrm{m}^{-3}$ ) was collected in February 2008 (Table 4).

At the first station, A. priodonta was dominant in the fall, winter and spring; Daphnia longispina was dominant in the winter and spring; M. gracilis was
dominant in the spring, summer and fall; Leptodora kindtii was dominant in the summer; Acanthocyclops robustus was dominant in the summer and spring.

At the second station, A. priodonta was dominant in the fall, winter and spring; B. longirostris, C. vicinus and $D$. galeata were dominant throughout the year; $D$. longispina was subdominant in the fall, winter and

Table 2. The number and density of the identified species (species and individuals per $\mathrm{M}^{3}$ ) at the first station of the Çaygören Reservoir

|  | Fb. 07 | Mr. 07 | Ap. 07 | My. 07 | Jn. 07 | J. 07 | Ag. 07 | Sp. 07 | Oc. 07 | Nv. 07 | Dc. 07 | Jn. 08 | Fb. 08 | Mr. 08 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of Species | 1 | 1 | 6 | 5 | 4 | 3 | 4 | 8 | 7 | 2 | 3 | 2 | 1 | 4 |
| Rotifera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Asplanchna priodonta | 0 | 0 | 4704 | 0 | 0 | 0 | 0 | 860 | 715 | 13780 | 1139 | 588 | 0 | 12576 |
| Gosse, 1850 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Pompholyx sulcata | 0 | 0 | 0 | 147 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hudson, 1885 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Keratella cochlearis (Gosse, 1851) | 0 | 0 | 784 | 0 | 0 | 0 | 122 | 214 | 0 | 0 | 0 | 0 | 0 | 0 |
| vulgaris Carlin, 1843 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| calyciflorus |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Pallas, 1766 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Filinia | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16440 | 286 | 0 | 0 | 0 | 0 | 0 |
| longiseta (Ehrenberg, 1834) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Trichocerca | 0 | 0 | 0 | 0 | 25948 | 0 | 3214 | 251 | 143 | 0 | 0 | 0 | 0 | 0 |
| (Wierzejski \& |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Zacharias, |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cladocera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Daphnia longispina | longispina |  |  |  |  |  |  |  |  |  |  |  |  |  |
| O.F.Müller, |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Leptodora | 0 | 0 | 0 | 12 | 1234 | 101 | 13 | 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| kindtii (Focke, 1844) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Copepoda |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cyclops vicinus | 674 | 10403 | 1568 | 294 | 0 | 0 | 0 | 234 | 2858 | 10963 | 1286 | 2352 | 172375 | 2401 |
| Uljanin, 1875 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Acanthocyclops robustus | 0 | 0 | 4312 | 1423 | 1533 | 3583 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| (G.O.Sars, 1863 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Metacyclops gracilis | 0 | 0 | 3920 | 11523 | 11616 | 24497 | 653 | 216 | 357 | 0 | 0 | 0 | 0 | 0 |
| (Lilljeborg, |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Eucyclops | 0 | 0 | 0 | 0 | 0 | 0 | 41 | 0 | 143 | 0 | 0 | 0 | 0 | 57 |
| speratus |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| (Lilljeborg, |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1901) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3. The number and density of the identified species (species and individuals per $M^{3}$ ) at the second station of the Çaygören Reservoir

|  | Fb. 07 | Mr. 07 | Ap. 07 | My. 07 | Jn. 07 | J. 07 | Ag. 07 | Sp. 07 | Ot. 07 | No. 07 | Dc. 07 | Jn. 08 | Fb. 08 | Mr. 08 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of species | 2 | 5 | 3 | 5 | 9 | 6 | 9 | 12 | 5 | 6 | 4 | 5 | 6 | 4 |
| Rotifera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Asplanchna priodonta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 860 | 0 | 13833 | 196 | 286 | 719 | 3528 |
| Gosse, 1850 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Pompholyx sulcata | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 359 | 0 |
| Hudson, 1885 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Keratella cochlearis (Gosse, 1851) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 214 |  | 0 | 0 | 0 | 0 | 0 |
| Polyarthra vulgaris Carlin, 1843 | 0 | 0 | 0 | 0 | 539 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notholca squamula (Müller, 1786) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 180 | 0 |
| Brachionus calyciflorus | 0 | 0 | 0 | 0 | 0 | 0 | 98 | 120 |  | 0 | 0 | 163 | 0 | 0 |
| Pallas, 1766 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Filinia <br> longiseta <br> (Ehrenberg, <br> 1834) | 0 | 0 | 0 | 0 | 0 | 0 | 19402 | 1644 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichocerca capucina (Wierzejski \& | 0 | 0 | 0 | 0 | 10150 | 0 | 24987 | 251 | 0 | 0 | 0 | 0 | 0 | 0 |
| $\begin{aligned} & \text { Zacharias, } \\ & \text { 1893) } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cladocera |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Daphnia longispina | 0 | 221 | 0 | 564 | 0 | 0 | 0 | 0 | 898 | 126 | 1254 | 998 | 198 | 653 |
| O.F.Müller, 1875 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Daphnia galeata Sars, 1864 | 0 | 490 | 0 | 4557 | 7096 | 5732 | 4018 | 422 |  | 16168 | 2090 | 13555 | 1617 | 2090 |
| Ceriodaphnia pulchella Sars, | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2450 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1862 <br> Bosmina | 0 | 817 |  | 2646 |  | 147 | 257 | 120 | 3185 | 2156 | 131 | 939 | 180 | 2678 |
| longirostris (O.F.Müller, 1785) | 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Moina micrura Kurz, 1874 | 0 | 0 | 0 | 0 | 269 |  | 3723 | 3150 | 2205 |  | 0 | 0 | 0 | 0 |
| Diaphanosoma brachyurum (Lievin, 1848) | 0 | 0 | 0 | 0 | 2605 | 4850 | 4997 | 1874 | 327 | 359 | 0 | 0 | 0 | 0 |
| Leptodora kindtii (Focke, 1844) | 0 | 0 | 0 | 1176 | 449 | 588 | 588 | 216 | 0 | 0 | 0 | 0 | 0 | 0 |
| Copepoda |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Cyclops vicinus Uljanin, 1875 | 0 | 0 | 98 | 0 | 3266 | 4491 | 1372 | 898 | 6341 | 9015 | 0 | 0 | 0 | 0 |
| Acanthocyclops robustus (G.O.Sars, 1863) | 898 | 4263 | 98 | 0 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Metacyclops <br> gracilis <br> (Lilljeborg, 1853) | 10779 | 9995 | 10485 | 294 | 1552 | 539 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 4. The number and density of the identified species (species and individuals per $\mathrm{M}^{3}$ ) at the third station of the Çaygören Reservoir

|  | Fb. 07 | Mr. 07 | Ap. 07 | My. 07 | Jn. 07 | J. 07 | Ag. 07 | Sp. | Oc. 07 | Nv. 07 | Dc. 07 | Jn. 08 | Fb. 08 | Mr. 08 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of species | 5 | 5 | - | 5 | 8 | 8 | 11 | 11 | 10 | 7 | 5 | 5 | 5 | 6 |
| Rotifera Asplanchna priodonta | 0 | 0 | - | 0 | 0 | 0 | 0 | 73 | 257 | 5063 | 102 | 225 | 784 | 245 |
| Gosse, 1850 <br> Pompholyx sulcata | 0 | 0 | - | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 392 | 0 |
| Hudson, 1885 <br> Keratella <br> cochlearis <br> (Gosse, 1851) | 0 | 0 | - | 0 | 0 | 0 | 97 | 110 | 0 | 0 | 0 | 0 | 0 | 0 |
| Keratella quadrata (Müller, 1786) | 0 | 0 | - | 0 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Brachionus calyciflorus Pallas, 1766 | 0 | 0 | - | 0 | 0 | 0 | 0 | 37 | 0 | 0 | 0 | 0 | 0 | 0 |
| Filinia longiseta (Ehrenberg, 1834) | 0 | 0 | - | 0 | 0 | 49 | 124 | 14294 | 0 | 0 | 0 | 0 | 0 | 0 |
| Trichocerca capucina (Wierzejski \& Zacharias, 1893) | 0 | 0 | - | 0 | 245 | 98 | 152 | 331 | 294 | 0 | 0 | 0 | 0 | 0 |
| Cladocera <br> Daphnia <br> longispina <br> O.F.Müller, <br> 1875 | 387 | 367 | - | 367 | 0 | 0 | 0 | 0 | 0 | 453 | 1450 | 812 | 149 | 490 |
| Daphnia galeata (Sars, 1890) | 645 | 1764 | - | 1764 | 1715 | 3675 | 2154 | 220 | 478 | 5226 | 296 | 9252 | 5683 | 1715 |
| Bosmina longirostris (Müller, 1776) | 122 | 588 | - | 588 | 82 | 73 | 169 | 73 | 331 | 1796 | 102 | 584 |  | 2205 |
| Moina micrura (Kurz, 1874) | 0 | 0 | - | 0 | 0 | 0 | 917 | 3087 | 588 | 0 | 0 | 0 | 0 | 0 |
| Diaphanosoma brachyurum (Lievin, 1848) | 0 | 0 | - | 0 | 286 | 3919 | 678 | 478 | 1139 | 327 | 0 | 0 | 0 | 0 |
| Leptodora kindtii (Focke, 1844) | 0 | 0 | - | 0 | 122 | 539 | 194 | 0 | 331 | 0 | 0 | 0 | 0 | 0 |
| Copepoda <br> Cyclops vicinus Uljanin, 1875 | 16903 | 3013 | - | 3013 | 0 | 0 | 216 | 73 | 1874 | 10779 | 1062 | 494 | 67026 | 10534 |
| Acanthocyclops robustus (G.O.Sars, 1863) | 0 | 0 | - | 0 | 531 | 931 | 349 |  | 257 | 490 | 0 | 0 | 0 | 0 |
| Metacyclops gracilis (Lilljeborg, 1853) | 94 | 220 | - | 220 | 3920 | 3332 | 993 | 37 | 992 | 0 | 0 | 0 | 0 | 0 |
| Eucyclops <br> speratus <br> (Lilljeborg, 1901) | 0 | 0 | - | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 245 |

spring; M. gracilis was dominant in the spring, summer and fall.

At the third station, A. priodonta was dominant in the spring, summer and fall; $C$. vicinus was dominant in the spring, fall and winter; $B$. longirostris was dominant the in spring; $D$. galeata was dominant throughout the year; $D$. longispina was dominant in the winter; $M$. gracilis was dominant in fall and summer; $A$. robustus was dominant in the summer.

The ANOVA results showed that the number of species were significantly different among the seasons ( $\mathrm{F}=1.7, \mathrm{P}<0.05$ ), but the stations ( $\mathrm{F}=0.76, \mathrm{P}>0.05$ ). The number of individual were significantly different among the seasons ( $\mathrm{F}=1.9, \mathrm{P}<0.05$ ) and stations ( $\mathrm{F}=2, \mathrm{P}<0.05$ ).

CCA showed that Keratella cochlearis was correlated to chl-a; C. vicinus was to water temperature; $D$. galeata and $D$. longispina to $\mathrm{NO}_{3}-\mathrm{N} ; \mathrm{M}$. gracilis and $A$. robustus to $\mathrm{PO}_{4}$ and the other dominant species were not correlated to any measured physicochemical parameters (Figure 2).

## Discussion

Based on Secchi disk depth (average1 m) and chla concentrations (average $17 \mu \mathrm{~g} \mathrm{l}^{-1}$ ), the Çaygören Reservoir can be classified as eutrophic (OECD, 1982).

Secchi disk depth was negatively correlated with nutrient concentrations at all stations. This was probably due to as nutrient levels rise, the abundance of phytoplankton increase, which absorb light causing reduced water transparency (LaBounty, 2008).

There was a positive correlation between conductivity and nutrient concentrations. Conductivity is often considered as parameter showing the degree of nutrient loading (Parinet, Lhote, \& Legube, 2004). Intensive agriculture has been practiced in surroundings of the Çaygören Reservoir, resulting in high nutrient loading and conductivity in the reservoir as agricultural nonpoint sources are a major contributing factor to surface water eutrophication worldwide.
A. priodonta was dominant during fall, winter and spring at all stations. This is an omnivorous species and common in temperate eutrophic lakes. It is an opportunistic filter feeder capable of influencing the phytoplankton community structure in lakes and reservoirs by taking advantage of the abundance of different groups during different seasons (Oganjan, Virro, \& Lauringso, 2013)
L. kindtii was dominant in the summer at the first station. This is a common invertebrate predator in temperate lakes (Herzing \& Auer, 1990). L. kindtii


Figure 2. The diagram of Canonical Correspondence Analysis (CCA) showing the relationships between the physicochemical parameters and the dominant zooplankton species. Abbreviations: Trichoce, Trichocerca capucina; Leptodor, Leptodora kindtii; Metacycl, Metacyclops gracilis; Diaphano, Diaphanosoma brachyurum; Daph I, Daphnia longispina; Daph g, Daphnia galeata; Acanthoc, Acanthocyclops robustus; Keratell, Keratella cochlearis; Bosmina, Bosmina longirostris; Brachion, Brachionus calyciflorus; Moina mi, Moina micrura; Filinia, Filinia longiseta; Cyclops, Cyclops vicinus.
selects its prey based on size, usually feeding on smaller cladocerans. This was probably why zooplankton community in the Çaygören Reservoir underwent changes in species composition from smallbodied species in spring to large-bodied species in the summer (Apaydın Yağcı \& Ustaoğlu, 2012; ApaydınYağcı et al., 2015).
B. longirostris, C. vicinus and D. galeata were dominant throughout the year at second and third stations. These species are common members of zooplankton community characteristic to temperate eutrophic lakes (Petrusek, Cerny, \& Audenaert, 2004; Bledzki \& Rybak, 2016). The Çaygören Reservoir is a eutrophic temperate reservoir providing suitable habitat for these cosmopolitan crustaceans.
B. longirostris is a small-bodied, filter-feeding cladoceran widely distributed throughout the world in all kinds of freshwater ecosystems regardless of their trophic state (Toth \& Kato, 1997). This study showed that $B$. longirostris was more tolerant than large bodied-cladocerans to environmental stress. Because 2007 and 2008 were the driest years in the last there decades in Turkey and the water level was extremely low in the Çaygören Reservoir.

CCA showed that D. longispina and D. galeata had high correlations to nitrate concentrations. This could be attributed to high abundance of Cyanobacteria in the summer. Elser \& Urabe (1999) noted that increased biomass of D. galeata when nitrogen concentrations increased in Lake Biwa, Japan. The phytoplankton of the Çaygören Reservoir is dominated by the nitrogen fixing Cyanobacteria during warm seasons (Çelik \& Sevindik, 2015).

Studies suggest that D. longispina, favor development of Cyanobacteria blooms by removing small edible phytoplankton (Leibold, 1989). High abundance of Cyanobacteria during the summer contributes to high nitrate concentrations by fixing nitrogen in the eutrophic water bodies (Presing, Herodek, Preston, \& Vörös, 2001). High abundance of nitrogen fixing Cyanobacteria during summer probably resulted in high correlation between D. longispina and $\mathrm{NO}_{3}$ in the Çaygören Reservoir.

CCA showed that $M$. gracilis had high correlations to phosphate concentrations. This was probably the result of a counteracting stimulus represented by low DO concentrations which cause the release of phosphorus in the sediment of the eutrophic reservoirs. M. gracilis was dominant during the summer months especially at the deep third station. Phosphorus is released from the sediments into the lake water in deep lakes in the anoxic hypolimnion during the stratification period (Sondergaard, Jensen, \& Jeppens, 2003).

CCA showed that $C$. vicinus was correlated to water temperature. C. vicinus is a cosmopolitan copepod that is common in Turkey's inland waters. In recent years, many studies reported high abundance of
the C. vicinus in Turkish reservoirs (Bekleyen, 2003; Saler \& Alış, 2014; ApaydınYağcı, Yılmaz, Yazıcıoğlu, \& Polat, 2015; Alış \& Saler, 2016; Saler, 2017). Although C. vicinus was common throughout the year at all stations, it had the highest abundance during the summer time. This explains the correlation between this species and high water temperature.
C. vicinus is known to feed on a variety of prey types such as algae, rotifers and small crustaceans, but it prefers rotifers when they are abundant (Devetter \& Seda, 2006). The year-long dominance of C. vicinus shows that it not only preys on rotifers, it can also use different source of food. Phytoplankton offers an alternative food source for this species when rotifers are scarce (Devetter \& Seda, 2003).

In summary, this is the first limnological study conducted on the Çaygören Reservoir and the study revealed that the trophic status of this reservoir was eutrophic (OECD, 1982). The annual average chlorophyll-a concentration of the reservoir is 18.25 $\mu \mathrm{gl}^{-1}$ and the annual average Secchi disk depth is 1.15 m. Karadzic, Subakov-Simic, Krizmanic, \& Natic (2010) discussed the range of physical and chemical parameters for the trophic classification of reservoirs in Serbia, based on their results, the Çaygören Reservoir can be considered as eutrophic. The dominant zooplankton species in the Çaygören Reservoir are cosmopolitan and characteristic to the eutrophic temperate lake zooplankton communities. Finally, the size selective predation by $L$. kindtii probably contributed to the dominance of the large-bodied zooplankton in this reservoir. When L. kindtii was abundant during the summer 2007, the zooplankton community underwent changes in species composition from small-bodied species in spring to large-bodied species in the summer.

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